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Molecular biosurveillance of wood-boring cerambycid beetles using DNA metabarcoding

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Abstract

Individual sorting and identification of thousands of insects collected in mass trapping biosurveillance programs is a labor intensive and time-consuming process. Metabarcoding, which allows for the simultaneous identification of multiple individuals in a single mixed sample, has the potential to expedite this process. However, detecting all the species present in a bulk sample can be challenging. In this study, we quantified the effectiveness of metabarcoding at detecting all species in six different mock communities of xylophagous cerambycid beetles. No significant differences in the number of species detected were observed between MinION, Illumina, and IonTorrent sequencing technologies. However, a greater number of individuals was detected and identified to species using MinION. In addition, the proportion of reads assigned to the species level was higher with Illumina technology. The three sequencing technologies also showed similar results in detecting and identifying closely related species and species at low abundance. The capture method greatly influences sample preservation and detection. Indeed, individuals captured using monopropylene and water had both lower DNA concentration and species detection rates compared to individuals killed using just an insecticide without any collection medium.

Keywords: Biological invasions, Biosecurity, *Cerambycidae*, Illumina®, IonTorrent®, Metabarcoding, Oxford Nanopore®, Xylophagous.

Introduction

Over the last few years, there has been an exponential increase in biological invasions that is expected to persist over the next decades (Seebens et al. 2021). This is primarily due to factors such as globalization, tourism, and global warming (Chown et al. 2015). Among the species introduced beyond their native range by human activities, insects are the most prevalent group (Seebens et al. 2018) and can cause a wide range of impacts. Non-native insects can affect native flora, fauna and ecosystems in various ways (Kenis et al. 2009) and can transmit pathogens and diseases, thus posing risks to public health (Mazza et al. 2014). Economic implications are also to be considered since many invasive insects are significant pests for agricultural crops and plantation forests, inducing huge costs for their management (Bradshaw et al. 2016).

Among these non-native insects, species associated with woody plants are more and more dominating, probably because of the growing trade of ornamental plants and wooden packaging material transported with international cargo shipments (Roques et al. 2016; Aukema et al. 2010). One of these important families of xylophagous beetles is the long-horned Cerambycidae, with more than 200 species affecting forestry, horticulture, and agriculture (Rossa and Goczał 2021) resulting in multimillion-dollar losses every year (Wang 2017). To detect potential new invasions of Cerambycids, biomonitoring programs have been set up over large geographical areas with intensive trapping campaigns extending over several years (Roques et al. 2023; Mas et al. 2023). However, rapidly evolving trades lead to changes in trade routes and imported goods which results in an increasing arrival of new non-native species. Many of these species have not been previously reported as invaders, some are not considered to be pests in their native ranges, and some could even be unknown to science (Seebens et al. 2018). As part of the European project HOMED (https://homed-project.eu/) 244 traps were set up across Europe (France, Italy, Spain, Switzerland, Portugal, Austria,

England, Greece, Slovenia, Netherlands, Bulgaria, Czech Republic, and in Sweden), 38 in Asia (China, Siberia, Russia), 11 in North America (USA, Canada), five in the Caribbean (Martinique) and four in Australia baited with generic lures, for simultaneous detection of multiple species (Roques et al. 2023). Thousands of captured cerambycids had to be sorted out and identified by expert taxonomists. This identification step is time-consuming and labor-intensive, thus limiting the rapid detection of non-native individuals among large numbers of native ones (Piper et al. 2019; Chua et al. 2023; Abeynayake et al. 2021). It is, however, essential that non-native species are identified as quickly as possible to allow their eradication before establishment and dispersal (Richardson et al. 2000; Blackburn et al. 2011; Giovani et al. 2020).

Traditional DNA barcoding, which allows taxonomic assignment of an individual based on the sequencing of a short fragment of the Cytochrome Oxidase 1 (COI) gene (658bp) (Hebert et al. 2003) has been successfully used to accurately identify cerambycid pest species for biomonitoring (Hodgetts et al. 2016, Wu et al. 2017, Kelnarova et al. 2019, Javal et al. 2021). Despite its numerous advantages, individual DNA barcoding remains a laborious and time-consuming approach in the context of mass trapped insects as it requires individual sorting of thousands of specimens, tissue sampling (often legs), extracting and amplifying DNA and finally sequencing each sample individually. But the recent application of high-throughput sequencing (HTS) technologies to DNA barcoding allows to expedite the production of thousands of DNA barcodes (deWaard et al. 2019; Srivathsan et al. 2021).

Metabarcoding is also based on high-throughput sequencing (HTS) technologies which generates a large number of short DNA sequences (reads), allowing the identification of multiple individuals simultaneously from a single mixed sample (hereafter called “bulks”) (Liu et al. 2020), such as all the individual insects captured in a single biomonitoring trap. sing DNA as a proxy for species detection and considering sequence variation within and among taxa, the metabarcoding approach is constrained by the completeness of the reference databases to accurately assign sequences to correctly identified taxa (Liu et al. 2020).

Although metabarcoding has several advantages, it still suffers from a number of methodological limitations that can make it unfitting for rapid biosecurity detection. In particular, the time required to process samples can be an issue (long delays between capturing individuals and obtaining sequencing results), particularly when the sampling sites are located far away from laboratories and transporting samples may require specific permits for certain species or when external providers are slow to sequence samples, which may hinder biomonitoring projects (Krehenwinkel et al. 2019; Egeter et al. 2022). Although MiSeq is generally recommended due to its lower error rate and well-established bioinformatic procedures, but Braukmann et al. (2019) demonstrated similar performance in sequence quality and insect species recovery compared to IonTorrent platforms (Ion Torrent PGM, and Ion Torrent S5).

In recent years, Oxford Nanopore Technologies® have released a portable sequencing platform, the MinION. This small sequencer can be connected via USB to a laptop to perform sequencing (Krehenwinkel et al. 2019) and the all-in-one version (Mk1C) even includes a screen and a computer in a portable format. The use of the MinION for a metabarcoding application offers the possibility of performing DNA sequencing of bulk samples directly on site without the need for transport or relying on external sequencing providers.

The main aim of our study was to assess the use of metabarcoding for the biosurveillance of Cerambycid wood-boring beetles. We evaluate the capability of three sequencing technologies Minion Nanopore, Illumina and IonTorrent to differentiate closely related cerambycid species and detect species present at low abundances in trap samples. Additionally, we examine the effect of various factors on species detection, such as the collection types (dry versus wet methods), DNA sample quality/quantity, and primer pair selection.

Materials and Methods

Taxa sampling

Specimens used in our experiments originated from Europe (France, Greece, Portugal, Spain), China (Beijing and Zhejiang Province) and USA (Michigan) (**Tab. S1**). Nearly all of them were

captured as part of a worldwide trapping experiment using multi-funnel traps baited with a generic attractant blend including eight Cerambycid pheromones (see details of the blend composition and trapping methods in Roques et al. 2023). To test for the impact of the collection methods on DNA preservation, we selected specimens killed according to two different procedures (**Tabs S1, S2**). The “dry” procedure involved placing a section of mesh impregnated with α-cypermethrin insecticide (Storanet®, BASF Pflanzenschutz Deutschland, Germany) into the trap basins, of which the bottom had been replaced with a wire mesh to allow drainage and keep specimens dry. In the “wet” procedure, trap basins were filled monopropylene glycol (MPG) and water (H_2O) in a 50:50 ratio to act as a surfactant and preservative. The tested specimens were collected between summer 2018 and summer 2021, with the exception of two specimens that were hand-captured in 1987 and 2012, respectively (**Tabs S1, S2**). Following collection, the beetles were preserved in ethanol 95°C and stored in at -20°C, except for the two hand-collected specimens, which were dried and pinned in insect boxes.

Mock community construction and DNA extraction

Six mock communities were constructed as follows:

Test 1: Identifying closely-related species and the impact of the capture method.

To assess the efficiency of the different sequencing technologies and primers to differentiate between species, bulks 1 and 2 were composed of closely-related species (**Tab. S1**). Five of seven specimens in Bulk 1 were captured using the "dry" method, while the eight of nine of specimens in Bulk 2 were captured using the "wet" method. One (bulk 2) or two (bulk 1) exceptions (wet *versus* dry) condition were added to the bulks as controls. Two legs were collected from each individual and pooled to constitute the bulks. The whole set of legs was then ground using flame-sterilized metal pestles to limit the risk of contamination. DNA from the ground material was extracted using the Qiagen DNeasy Blood and Tissue Kit following the manufacturer’s instructions. Two other legs were taken from the specimens, from which we assessed the quantity and quality of DNA for each specimen individually (**Fig. 1a**).

Test 2: Detecting low abundance specimens.

Bulks 3 and 4 were composed of six species represented by heterogeneous DNA concentrations (**Tab. S1**) to assess the ability of the sequencing technologies and primers to detect species present in a very low abundance. The DNA of each individual was extracted using two legs that were ground as above and processed using the Qiagen DNeasy Blood and Tissue Kit. To construct bulks 3 and 4, individual DNA extracts were quantified using a fluorometer (Nanodrop™, Thermo Fisher Scientific) and mixed together according to their concentration to achieve the desired proportions of DNA for each individual (6 individuals of different species ranging from 41% to 3% for Bulk 3 and 6 individuals of different species ranging from 50% to 0.5% for bulk 4). All individuals in bulks 3 and 4 were captured using the "dry" collection method. (**Tab. S1, Fig. 1b**).

Test 3: Mimicking field trap content on species composition.

Bulks 5 and 6 comprised individuals from a number of species native to Europe usually found in the traps deployed there, with the addition of non-native species which have either already been introduced or are still not present in Europe (Bulk 5: 22 individuals of eight species, including two older specimens from 2012 and 1987 and including one non-native; Bulk 6: 41 individuals of 12 species including two non-native ones). Six individuals in bulk 5 were captured using the "dry" collection method and two individuals were hand-collected. Two specimens were captured using the "wet" method in bulk 6 (**Tab. S1**). The DNA was extracted following the same protocol as for bulks 1 and 2 (**Fig. 1a**).

PCR amplification

All bulk samples were amplified with the following two primer pairs: BF3/BR2 (call hereafter “B”) (CCHGAYATRGCHTTYCCHCG / TCDGGRTGNCCRAARAAYCA (Elbrecht and Leese 2017; Elbrecht et al. 2019), which generates a 458 bp amplicon; and fwhF2/fwhR2n (call hereafter “F”) (GGDACWGGWTGAACWGTWTAYCCHCC / GTRATWGCHCCDGCTARWACWGG), which generates a 254 bp amplicon (Vamos et al. 2017). Each PCR comprised 15.3 µl H2O, 2.5 µl 10X PCR buffer, 2.5 µl dNTP [1mM], 1 µl of each primer [0,4mM], 0.2 µl Dream Taq (Thermo Fisher Scientific), 0.5 µl Betaine [100mM], 2 µl DNA and H2O for a total of 25 µl per reaction. PCR was performed using the following program: 95°C for 5 min, 29 cycles of 95°C for 30 s, 48°C for 30

s, and 72°C for 50 s and 72°C for 5 min (Elbrecht et al. 2019). PCR products were then run on a 2% agarose gel stained with ethidium bromide and visualised by UV transluminator. The PCR products were then purified with the NucleoFast 96 PCR plate purification kit (Macherey-Nagel).

Illumina® library preparation

A second PCR was performed on the products of the first PCR to add Illumina® tags and adapters, prepared by ligating Nextera XT indices through an eight cycle PCR. The second PCR was carried out with the same conditions as for the initial PCR. Reactions comprised PCR amplification reactions (25 µl) contained the following: 2 µl of template DNA, 1.5 µl of each primer [10 µM], 5 µl of 5X GoTaq (Promega) reaction buffer, 1 µl of MgCl2 [25 mM}, 1 µl of BSA [1 mg/ml], 0.5 µl of dNTPs [5 mM], 13.87 µl of molecular-grade water and 0.13 µl of GoTaq G2 Polymerase (Promega), 5 µl of the purified products from the first PCR, and 2 µl of barcodes. The PCR conditions were the same as for the first PCR, with eight cycles. The products of the second PCR were verified on a 2% agarose gel. PCR products were then equimolarly pooled into two different pools (one pool per primer pair used) and purified using the GeneJET Gel Extraction kit on an agarose gel, following manufacturer's instructions. This library was sequenced in Illumina MiSeq using V2 chemistry (300 × 300 bp, 500 cycles) in the Sequencing Center within the Biozentrum of the Ludwig-Maximilian University in Munich (Germany).

MinION library preparation

Libraries were prepared according to the Oxford Nanopore Technologies ® protocol: "PCR barcoding (96) amplicons (SQK-LSK110) (version: PBAC96_9114_v110_revF_10Nov2020)" with the following specifications. The PCR barcoding expansion Pack 1-96 (EXP-PBC096) was used to perform the second PCR to incorporate the Oxford Nanopore Technologies ® barcode sequences on the amplicons generated in the first PCR. Final PCR products were then quantified using Qubit and equimolarly pooled before being purified with Agencourt AMPure XP beads (Beckam Coutler). The final pool was then sequenced on the MinION sequencer (Mk1c; Oxford Nanopore Technologies ®, UK) using a R10.3 flowcell (MIN111) with 1331 pores available and the LSK110 ligation sequencing kit.

IonTorrent® library preparation

For the production of the libraries, we started with 5 ng of DNA (Qubit measurement). The Nextflex Cellfree DNAseq kit (PerkinElmer) was used for the process. The quality of the libraries was assessed using Qubit (for quantification) and Bioanalyzer (using the HighSensitivity kit from Agilent, for size verification). After quality control, each library was amplified by emulsion PCR on the Ion One Touch 2 instrument, with a concentration of 15 pg/µl. Subsequently, the libraries were sequenced on an Ion GeneStudio S5 system using a single-end sequencing protocol with a 300 bp read length. Sequencing was performed on an Ion 520 Chip by the GeT-BioPuces platform (Toulouse, France).

Reference Barcode Dataset

A dataset was built using all the public sequences of Cerambycidae available in BOLD systems (Ratnasingham and Hebert 2007). It was then verified whether all 33 species present in the bulk samples were represented by at least one sequence in the database. We found that three species were not present in BOLD and we therefore barcoded them by Sanger sequencing on an ABI 3500 genetic analyzer (Applied Biosystems) using the big-dye terminator sequencing V3.0 kit (Applied Biosystems). The three newly generated barcodes along with one barcode per species represented in the mock communities are available from BOLD in the dataset DS-MINION (dx.doi.org/XX/ DS-MINION).

Illumina® data processing

The raw data was analysed using the *FROGS* pipeline, a standardized pipeline containing a set of tools that are used to process amplicon reads that have been produced from Illumina® sequencing (Escudié et al. 2018; Henrie et al. 2022). First, amplicons with a size between 408 and 508 for the BF3/BR2 primer pair and 204 and 304 for the fwhF2/fwhR2n primer pair were retained. Sequence clustering was then performed using the SWARM algorithm (Mahé et al. 2014) with a maximum sequence difference set at d=1, as recommended by SWARM. Chimeric sequences were then removed. Sequences were aligned to the same database used for the MinION and IonTorrent® data analysis. In order to remove all spurious detections, OTU

detections with less than 10 reads were removed. The identification was considered as 'valid' at species-level from a similarity threshold ≥ 98% (Alberdi et al. 2018). Below that threshold, OTUs were considered unidentified.

MinION and IonTorrent® data processing

Bioinformatics analyses were performed on the Genotoul Bioinformatics Platform (INRAE, Toulouse, France). Basecalling and demultiplexing were performed for MinION data using Guppy v6.1.7; ONT; high accuracy base calling mode; parameters: -c dna_r10.3_450bps_hac.cfg --min_qscore 5 --trim_barcodes. Then, for MinION and IonTorrent® data, we used the *msi* data processing pipeline v0.3.6 (Egeter et al. 2022) to reduce the error rate of the reads by polishing them after the basecalling step. Reads smaller than 40bp were removed with cutadapt v4.0 (Martin 2011). The size range was set between 408bp and 508bp for BF3/BR2 and between 204bp and 304bp for fwhF2/fwhR2n. The clustering step was carried out with *ISONCLUST* v0.0.6.1 (Sahlin and Medvedev 2020; with parameters: --mapped_threshold 0.825 and --aligned_threshold 0.55) and a consensus sequence per cluster was generated using RACON v1.5.0 (Vaser et al. 2017). The polished reads were then clustered at 97% sequence identity with *CD-HIT* v4.8.1 (Fu et al. 2012) and a representative sequence from each cluster (centroid) was selected. The polished reads were then aligned to the local database with BLAST. The following parameters were used: -word_size 11 -perc_identity 95 -qcov_hsp_perc 98 -gapopen 0 -gapextend 2 -reward 1 -penalty 1 -max_target_seqs 100. Finally, a taxonomic assignment was performed for each query using a lowest common ancestor (LCA) approach with the bioinformatics package metabinkit (Chain et al. 2016; Egeter et al. 2018 Kitson et al. 2019) with the following parameters: 98% at species level, 97% at genus level, 95% at family level (Alberdi et al. 2018; Egeter et al. 2022). Similarly, to the Illumina® data processing, OTU detections with less than 10 reads were removed.

Statistical analysis

The Wilcoxon test was used in R to assess the significance of the impact of the conservation method (dry or wet) on sample DNA concentration, sample DNA quality, and associated detection counts. A two-sample test of proportions was used to compare and assess the significance of the proportion of reads assigned to the species levels for MinION, Illumina, and

IonTorrent technologies using the "Social Science Statistics" website (https://www.socscistatistics.com/tests/anova/default2.aspx). The proportion of reads assigned to different taxonomic levels was calculated by summing the total reads from different bulk samples for each condition. To determine if the number of false positives was significantly different among the 3 technologies and the two primer pairs, we calculated the detection mean for each bulk under different conditions. We then performed an ANOVA test followed by a Tukey HSD test using the "Social Science Statistics" website.

Results

A total of 1.248.95 reads were sequenced for the MinION Nanopore® technology using the fwhF2/fwhR2n primer pair, with an average of 78.037 reads per sample. After quality filtering, removing reads of incorrect size or insufficient quality, 1.113.844 (89.2%) reads were retained, with an average of 69.615 reads per bulk. For the BF3/BR2 primer pair, a total of 1.132.604 reads were sequenced, with an average of 62.922 reads per sample. After quality filtering, a total of 948.832 (83.8%) reads were retained, with an average of 52.712 reads per bulk. The Illumina® sequencing produced a total of 1.549.894 reads using the BF3/BR2 primer pair, with an average of 258.316 reads per bulk. After quality filtering, 1.025.637 (66.2%) reads were retained, with an average of 170.940 reads per bulk. For the fwhF2/fwhR2n primer pair, a total of 2.299.072 reads were sequenced, with an average of 383.179 reads per bulk. After quality filtering, 1.686.058 (73.3%) reads were retained, with an average of 281.010 reads per bulk. For the IonTorrent® technology, 838.489 reads were sequenced, with an average of 139.748 reads per bulks with the BF3/BR2 primer pair. After the quality filtering, 280.695 (33.5%) reads remains with an average of 46.782 reads per bulks.

Benchmarking of sequencing technologies

The MinION technology accurately identified 28 out of 48 specimens at the species level, Illumina® technology allowed specific identification of 27 specimens and IonTorrent® identified 24 specimens. The primer pair fwhF2/fwhR2n allowed specific identification of 27 specimens with two OTU of its own while the primer pair BF3/BR2 allowed the identification of 31 specimens with six OTU of its own. Illumina® F, Illumina® B and MinION F allowed for 25

species-level identifications across all bulks, compared to 24 when considering the combination MinION F and IonTorrent® B. This difference was not significant. The number of identifications only obtained with one combination varied from zero (IonTorrent® B) to two (Minion B and Illumina® B) (**Fig. 2**).

The proportion of reads assigned at the species taxonomic level was significantly higher with Illumina® technology (p.value < 0.00001) comparing with the MinION, all primers included, particularly when considering primer pair F. Nearly 97% of reads were assigned at the species level for the Illumina® F combination compared to 90% for the MinION F combination (p.value<0.0001). As for primer pair BF3/BR2, over 87.3% of reads were assigned at the species level for Illumina®, followed by over 79.7% for MinION technology and over 77.2% for IonTorrent® technology (**Fig. 3**). The couple of primer fwhF2/fwhR2n resulted in a significantly higher percentage of reads assigned at the species level (93.6%) (considering both Illumina® and MinION technologies) compared to couple of primers B (81.4%) (considering all three technologies) (p.value<0.00001).

False positive detections were observed for each combination of primers and technology (**Fig. 4**). Hence, an average of 13.5 false positives OTU were recorded for the primer pair fwhF2/fwhR2n, compared to an average of 4 false positives OTU when using the primer pair BF3/BR2, the difference being significant here (p.value = 0.00194). According to the technology used, regardless of the primers, an average of 10 / 7 and 6 false positives were recorded for Illumina®/MinION/IonTorrent® technologies respectively. There are no significant differences among the three sequencing technologies in terms of false positives.

Comparative study within bulks

In total, 33 out of 48 individuals (68.8%) were detected at the species level by at least one experimental condition (**Fig. 5**).

The objective of Bulks 1 and 2 was to compare the detection rates of closely related species under different conditions. No significant differences were observed among the different

methods used. Illumina® detected seven species out of 16 (43.75%), MinION also detected seven out of 16 (43.75%), and IonTorrent® detected six species out of 16 (37.5%). However, significant differences were observed among the studied taxonomic groups: three out of four species from the genus *Arhopalus* were detected, as well as for *Xylotrechus*. Only two out of four species were detected for the genus *Monochamus*, and one out of five species for the genus *Phymatodes*.

Metabarcoding of bulks 3 and 4 aimed at comparing the ability of different sequencing technologies to detect species present in low abundance in the traps. All sequencing technology/primer combinations allowed for the detection of minor species in bulks 3 and 4: *Phymatodes testaceus* with a presence of 3% in bulk 3 and *Xylotrechus chinensis* with a percentage of 0.5% in bulk 4. However, some species (although not in minority in the bulks) were not detected in one or several test (**Fig. 5**). In total, Illumina® was able to detect a higher number of individuals (11 out of 12 individuals detected) compared to MinION (nine of 12 individuals detected) and IonTorrent® (nine of 12 individuals detected).

In bulk 5, the non-native species, *Cordylomera spinicornis* was detected only by Illumina B. For bulk 6, the non-native species *Xylotrechus chinensis* was detected by all conditions and *Xylotrechus stebbingi* by MinION B only. The results showed that MinION performs better to detect and identify trapped species (detecting eight out of 12 species (66.7%)) compared to Illumina® and IonTorrent® technologies (which detected five out of 12 species (41.7%)) for Bulk 6. As for Bulk 5, all technologies detected the same number of species (four out of six (66.7%)).

Impact of capture and storage conditions on individual detection

Our results demonstrate a significative difference in the mean number of detections between samples that were collected using the “dry” method (α-cypermethrin insecticide) and the “wet'” method (water-diluted propylene glycol) (Wilcoxon rank-sum test, W = 74.5, p.value = 0.0006342) (**Fig. 6**). Almost all specimens (9 out of 12 - 75%) collected using the “wet” trapping procedure (water-diluted propylene glycol) could not be detected by any of the sequencing technologies, including MinION. Conversely, those collected using the “dry” trapping

procedure (using α-cypermethrin insecticide) had 30 out of 34 detected specimens (88.2%), by each of the sequencing technologies.

Regarding the impact of the collection type on the concentration and quality of DNA samples, individuals captured using the "dry" method had an average DNA concentration of 39 ng/µl and an average DNA concentration of 18.6 ng.µl for the “wet” method (Wilcoxon rank-sum test, W = 123.5, p.value = 0.04533). We observe an average A260/280 ratio of 1.9 for the “dry” method and an average of 2 for “wet” method. Finally, we observe an average A260/230 ratio of 0.8 for the “dry” method, while samples captured using the "wet" approach had an average A260/230 ratio of 0.5 (Wilcoxon rank-sum test, W = 146, p.value = 0.1502) (**Fig. 6**).

The specimens collected by hand and kept pinned dry were not detected by any of the sequencing technologies. Species with a "dry" collection type trapped in 2021 have a detection rate of 80% (detected by at least one condition), those trapped in 2020 have a detection rate of 86.7%, and those from 2019 have a detection rate of 100%. Both, individuals captured in 2018 and 2010 were detected, unlike those captured in 1987 and 2012.

Discussion

Benchmarking of sequencing technologies

No significant differences in the number of species detected were observed between the three sequencing technologies, even if a greater number of individuals was detected and identified to species using MinION (28 specimens) compared to Illumina® (27 specimens) or IonTorrent® (24 specimens). These results are consistent with the findings of Srivathsan et al. (2021), which demonstrated that MinION barcodes are nearly identical to Sanger and Illumina barcodes for the same samples. It must be taken into account that we worked on a single pair of primers with the IonTorrent® technology, which may have reduced the number of detections. Detection rate was higher when using the primer pair generating a larger amplicon size, BF3/BR2. In addition, only BF3/BR2 allowed the species-level identification of *Xylotrechus stebbingi*. This difference may be due to the longer amplicon generated by this primer pair

(458bp), which has more genetic information and therefore more data to provide a reliable taxonomic assignment.

The Illumina® technology has a higher percentage of reads allowing species-level identification compared to MinION or IonTorrent®. The detection of specimens at only higher taxonomic level (genus or family), can be explained by sequencing errors that produce reads with less than 98% identity to the reference database. Thus our results suggest that Illumina® generates slightly less sequencing errors than MinION and IonTorrent®. It is also noted that the primer pair fwhF2/fwhR2n produced a higher percentage of reads allowing species-level identification than the primer pair BF3/BR2. However, as seen previously, fwhF2/fwhR2n generated a significant higher number of false positives than BF3/BR2 (**Fig. 4**). This may be due to the fact that fwhF2/fwhR2n is smaller in size than BF3/BR2 thus any loss of genetic information is more likely to lead to misidentification or false negatives.

The three technologies also showed similar results in detecting and identifying closely related species. Moreover, the results show that all three sequencing technologies (regardless of the associated primer pairs) enabled the detection and identification of species whose DNA represented a very low percentage of the DNA extraction (**Fig. 5**). Thus, all three technologies appear suitable for detecting and identifying species in low number in traps.

Impacts of capture and storage conditions on DNA conservation

Our results show an impact of sample capture conditions. Individuals captured using the “wet” method had lower DNA concentration and presented significantly much lower detection rates compared to individuals captured using the “dry” method. Ballare *et al.* (2019) also found that insects collected in propylene glycol traps produced lower-quality ddRADseq assemblages. On the contrary, Ferro and Park, 2013 reported that propylene glycol is an effective DNA preservative for molecular marker-based studies. However, in mentioned study, the insects were initially killed and preserved in 100% ethanol before being stored to glycol, unlike our study where the insects were initially killed in propylene glycol. The use of 100% ethanol as the initial agent for destruction and preservation may have resulted in better initial

preservation of specimens than if the specimens had been directly exposed to propylene glycol.

The conservation method may also play a role in species detection, as the two specimens conserved using the “pinned dry” method were not detected. However, we have insufficient “pinned dry” specimens in this study to confirm this hypothesis. But numerous studies have already demonstrated that it is possible to use dry insect specimens for genetic analyses, although such types of analyses are much more complex than when using fresh specimens (Wandeler et al. 2007; Hebert et al. 2013; Nakahama et al. 2018).

The storage duration does not appear to significantly impact species detection. We did not observe a decrease in detection rates between individuals collected after 2018. However, the impact of storage duration seems more pronounced for older individuals (collected between 1987 and 2012). Once again, we have limited data to draw solid conclusions regarding these older samples although it is known that the storage time plays an important role in DNA yield, fragment size, and PCR success (Dean and Ballard 2001).

Ultimately, the DNA concentration of the samples appears to play an important role in their detection as DNA extracts with higher concentration were more likely to be detected (DNA concentration of 38.6 ng/µl for detected samples compared to 19.6 ng/µl for undetected samples). The quality ratios A260/230 and A260/280 are similar between detected and undetected samples, suggesting they do not contribute to the observed detection rates in our analysis.

False positives, negatives and unmatched OTUs

Despite precautions taken, several false positives were detected for all tested conditions. The number of false positives is significantly higher with the primer pair fwhF2/fwhR2n, which is smaller in size, potentially leading to incorrect taxonomic identifications compared to BF3/BR2, which is larger in size. Despite the fact that Illumina technology is known to have a lower sequencing error rate compared to MinION, our study found ten false positives

generated by Illumina, while MinION had seven false positives and IonTorrent had six false positives. The sensitivity of HTS technologies allows for the detection of very small amounts of DNA, thus detecting even the slightest cross contamination between samples (Liu et al. 2020). These contaminations may have occurred during sample collection in the field or in the laboratory through cross-contamination between samples from the same study.

The false negative detections for a certain number of individuals can primarily be explained by the highly heterogeneous DNA quality of the different sequenced individuals (**Tab. S1**). In fact, DNA quality can be impacted by numerous mainly abiotic factors (pH, UV radiation, temperature), degrading DNA quality in a matter of days/weeks (Strickler et al. 2015; Collins et al. 2018; Harrison et al. 2019). During field trapping using stationary traps, captured insects are sometimes exposed to such conditions (high temperatures in trap containers when exposed to the sun in summer, high humidity in the container during heavy rains, ...), which can greatly accelerate the speed of DNA degradation in captured individuals. This degraded DNA is more difficult to be amplified, thus generating false negatives, especially when attempting to detect low-abundance insects in the trap, such as an invasive species in the process of establishing (Preston et al. 2022). Another possible cause for the high number of false negatives is the bias induced by PCR, such as uneven amplification of the DNA of the different individuals present in one sample (Preston et al. 2022). To avoid potential bias arising from identification mistakes due to errors or missing species in the references databases, we decided to work on a local and curated BLAST database. However, when target species are partially unknown, as is the case in field conditions, analyses must rely on public reference databases. Yet, out of the 35,000 known species of Cerambycidae to date, only 2,884 species (8%) are recorded in BOLD with a barcode fragment (as of April 17th, 2023). Furthermore, as mentioned above, databases can contain errors such as misassignment of an DNA sequence to a wrong species (due to morphological identification errors). This was precisely the error encountered for the species *Monochamus sutor* who has been identified as *Monochamus sartor* or the species *Leiopus nebulosus* who has been identified as *Leiopus linnei*.

Morphological similarities or identifications from non-specialists can lead to errors in databases, hindering their identification at species level. One also needs to pay attention to

synonymy when a species is called by multiple names. We encountered this problem in our analysis where *Arhopalus ferus* (Bulks 1, 3, and 6) was detected but under the name of *Arhopalus tristis*. Finally, mitochondrial paralogues such as Numts (non-functional copies of mitochondrial genes transported into the nuclear genome) present in databases can also bias results, making it impossible to identify correctly the species concerned (Bensasson et al. 2001). Numts are numerous in many organisms, including some cerambycids such as *Monochamus galloprovincialis (*Koutroumpa et al. 2009; Haran et al. 2015).

Biaises

Based on the results obtained, it appears that the main biases observed during metabarcoding analyses on trap contents come from the degradation of DNA from individuals, which generates false negatives. To limit the biases induced by the degradation of DNA samples, it is important, when possible, to favour a "dry" rather than a "wet" trap and to plan for the collection, transportation, and processing of captured individuals. This includes regularly checking the traps as frequently as possible, thus avoiding excessively long exposure of the individuals to unfavourable environmental conditions. Once individuals are brought back to the laboratory and if DNA cannot be extracted straight away, it is important to limit any further degradation by keeping samples in a -20°C freezer in 95% ethanol. DNA extractions, on the other hand, should be stored in the preservation buffer provided with the extraction kits and kept at -20°C (Preston et al. 2022). We also recommend limiting the use of primer pairs that generate short amplicons, which can favour the amplification of non-target taxa and lead to identification errors. The quality and completeness of the databases is also a very important bias factor. To limit this bias, it is recommended to restrict the database used to targeted species in order to minimize the risk of false positives due to contamination (Egeter et al. 2022). Limited taxonomic and geographical coverage of sequence databases is a huge limitation in metabarcoding studies (for example, Dopheide et al. (2019) found no representative sequence in the GenBank database for more than 900 invertebrate OTUs in one study). Additionally, identification errors of species and cases of synonymy lead to false negatives or cases of multiple affiliations.

Conclusion

By comparing the accuracy and detection capacity of three metabarcoding strategies, this study contributes to improve our toolkit for insect invasion monitoring. All three sequencing technologies performed equally well and showed similar results for the detection and identification of invasives species in the traps, but as a portable, easy-to-use and cost-effective sequencer, the MinION has the potential to become an essential tool for biodiversity monitoring projects. Indeed, using the MinION saves laboratory handling time compared with Illumina and avoids outsourcing sample sequencing, saving considerable time. This technology is precise enough to detect species present at low abundances in traps and allows for accurate identifications as long as there is a sufficiently high-quality reference database to avoid identification errors or false positives/negatives. It is also crucial to pay close attention to issues of contamination and insect preservation during and after individual capture to work with the least degraded DNA possible.

References

Abeynayake SW, Fiorito S, Dinsdale A, Whattam M, Crowe B, Sparks K, Campbell PR, Gambley C (2021) A Rapid and Cost-Effective Identification of Invertebrate Pests at the Borders Using MinION Sequencing of DNA Barcodes. Genes 12: 1138. https://doi.org/10.3390/genes12081138

Alberdi A, Aizpurua O, Gilbert MTP, Bohmann K (2018) Scrutinizing key steps for reliable metabarcoding of environmental samples. Mahon A (Ed.). Methods in Ecology and Evolution 9: 134–147. https://doi.org/10.1111/2041-210X.12849

Aukema JE, McCullough DG, Von Holle B, Liebhold AM, Britton K, Frankel SJ (2010) Historical Accumulation of Nonindigenous Forest Pests in the Continental United States. BioScience 60: 886–897. https://doi.org/10.1525/bio.2010.60.11.5

Ballare KM, Pope NS, Castilla AR, Cusser S, Metz RP, Jha S (2019) Utilizing field collected insects for next generation sequencing: Effects of sampling, storage, and DNA extraction methods. Ecology and Evolution 9: 13690–13705. https://doi.org/10.1002/ece3.5756

Bensasson D, Zhang D-X, Hartl DL, Hewitt GM (2001) Mitochondrial pseudogenes: evolution's misplaced witnesses. Trends in Ecology & Evolution 16: 314–321. https://doi.org/10.1016/s0169-5347(01)02151-6

Blackburn TM, Pyšek P, Bacher S, Carlton JT, Duncan RP, Jarošík V, Wilson JRU, Richardson DM (2011) A proposed unified framework for biological invasions. Trends in Ecology & Evolution 26: 333–339. https://doi.org/10.1016/j.tree.2011.03.023

Bradshaw CJA, Leroy B, Bellard C, Roiz D, Albert C, Fournier A, Barbet-Massin M, Salles J-M, Simard F, Courchamp F (2016) Massive yet grossly underestimated global costs of invasive insects. Nature Communications 7: 12986. https://doi.org/10.1038/ncomms12986

Braukmann TWA, Ivanova NV, Prosser SWJ, Elbrecht V, Steinke D, Ratnasingham S, De Waard JR, Sones JE, Zakharov EV, Hebert PDN (2019) Metabarcoding a diverse arthropod mock community. Molecular Ecology Resources 19: 711–727. https://doi.org/10.1111/1755-0998.13008

Chain FJJ, Brown EA, MacIsaac HJ, Cristescu ME (2016) Metabarcoding reveals strong spatial structure and temporal turnover of zooplankton communities among marine and freshwater ports. Cowie R (Ed.). Diversity and Distributions 22: 493–504. https://doi.org/10.1111/ddi.12427

Chown SL, Hodgins KA, Griffin PC, Oakeshott JG, Byrne M, Hoffmann AA (2015) Biological invasions, climate change and genomics. Evolutionary Applications 8: 23–46. https://doi.org/10.1111/eva.12234

Chua PYS, Bourlat SJ, Ferguson C, Korlevic P, Zhao L, Ekrem T, Meier R, Lawniczak MKN (2023) Future of DNA-based insect monitoring. Trends in Genetics 39: 531–544. https://doi.org/10.1016/j.tig.2023.02.012

Collins RA, Wangensteen OS, O'Gorman EJ, Mariani S, Sims DW, Genner MJ (2018) Persistence of environmental DNA in marine systems. Communications Biology 1: 185. https://doi.org/10.1038/s42003-018-0192-6

Dean MD, Ballard JWO (2001) Factors affecting mitochondrial DNA quality from museum preserved Drosophila simulans. Entomologia Experimentalis et Applicata 98: 279–283. https://doi.org/10.1046/j.1570-7458.2001.00784.x

deWaard JR, Levesque-Beaudin V, deWaard SL, Ivanova NV, McKeown JTA, Miskie R, Naik S, Perez KHJ, Ratnasingham S, Sobel CN, Sones JE, Steinke C, Telfer AC, Young AD, Young MR, Zakharov EV, Hebert PDN (2019) Expedited assessment of terrestrial arthropod diversity by coupling Malaise traps with DNA barcoding 1. Genome 62: 85–95. https://doi.org/10.1139/gen-2018-0093

Dopheide A, Tooman LK, Grosser S, Agabiti B, Rhode B, Xie D, Stevens MI, Nelson N, Buckley TR, Drummond AJ, Newcomb RD (2019) Estimating the biodiversity of terrestrial invertebrates on a forested island using DNA barcodes and metabarcoding data. Ecological Applications 29: 1–14. Available from: https://www.jstor.org/stable/26669248 (June 9, 2023).

Egeter B, Peixoto S, Brito JC, Jarman S, Puppo P, Velo-Antón G (2018) Challenges for assessing vertebrate diversity in turbid Saharan water-bodies using environmental DNA. Genome 61: 807–814. https://doi.org/10.1139/gen-2018-0071

Egeter B, Veríssimo J, Lopes-Lima M, Chaves C, Pinto J, Riccardi N, Beja P, Fonseca NA (2022) Speeding up the detection of invasive bivalve species using environmental DNA: A Nanopore

and Illumina sequencing comparison. Molecular Ecology Resources 22: 2232–2247. https://doi.org/10.1111/1755-0998.13610

Elbrecht V, Leese F (2017) Validation and Development of COI Metabarcoding Primers for Freshwater Macroinvertebrate Bioassessment. Frontiers in Environmental Science 5. Available from: https://www.frontiersin.org/articles/10.3389/fenvs.2017.00011 (July 2, 2023).

Elbrecht V, Braukmann TWA, Ivanova NV, Prosser SWJ, Hajibabaei M, Wright M, Zakharov EV, Hebert PDN, Steinke D (2019) Validation of COI metabarcoding primers for terrestrial arthropods. PeerJ 7: e7745. https://doi.org/10.7717/peerj.7745

Escudié F, Auer L, Bernard M, Mariadassou M, Cauquil L, Vidal K, Maman S, Hernandez-Raquet G, Combes S, Pascal G (2018) FROGS: Find, Rapidly, OTUs with Galaxy Solution. Berger B (Ed.). Bioinformatics 34: 1287–1294. https://doi.org/10.1093/bioinformatics/btx791

Ferro ML, Park J-S (2013) Effect of Propylene Glycol Concentration on Mid-Term DNA Preservation of Coleoptera. In: The Coleopterists Bulletin. , 581–586. https://doi.org/10.1649/0010-065X-67.4.581

Fu L, Niu B, Zhu Z, Wu S, Li W (2012) CD-HIT: accelerated for clustering the next-generation sequencing data. Bioinformatics (Oxford, England) 28: 3150–3152. https://doi.org/10.1093/bioinformatics/bts565

Giovani B, Blümel S, Lopian R, Teulon D, Bloem S, Galeano Martínez C, Beltrán Montoya C, Urias Morales CR, Dharmapuri S, Timote V, Horn N, Chouibani M, Mezui M'Ella JG, Herrera V, Castinel A, Goletsos C, Moeller C, Naumann I, Stancanelli G, Bronzwaer S, Tramontini S, MacDonald P, Matheson L, Anthoine G, De Jonghe K, Schenk M, Steinmöller S, Rodriguez E, Cruz ML, Luck J, Fraser G, Brunel S, Montuori M, Fedchock C, Steel E, Pennington HG, Day R, Rossi JP, Xia J (2020) Science diplomacy for plant health. Nature Plants 6: 902–905. https://doi.org/10.1038/s41477-020-0744-x

Haran J, Koutroumpa F, Magnoux E, Roques A, Roux G (2015) Ghost mtDNA haplotypes generated by fortuitous NUMTs can deeply disturb infra-specific genetic diversity and phylogeographic pattern. Journal of Zoological Systematics and Evolutionary Research 53: 109–115. https://doi.org/10.1111/jzs.12095

Harrison JB, Sunday JM, Rogers SM (2019) Predicting the fate of eDNA in the environment and implications for studying biodiversity. Proceedings of the Royal Society B: Biological Sciences 286: 20191409. https://doi.org/10.1098/rspb.2019.1409

Hebert PDN, Ratnasingham S, deWaard JR (2003) Barcoding animal life: cytochrome c oxidase subunit 1 divergences among closely related species. Proceedings of the Royal Society B: Biological Sciences 270: S96–S99. https://doi.org/10.1098/rsbl.2003.0025

Hebert PDN, deWaard JR, Zakharov EV, Prosser SWJ, Sones JE, McKeown JTA, Mantle B, La Salle J (2013) A DNA 'Barcode Blitz': Rapid Digitization and Sequencing of a Natural History Collection. Kolokotronis S-O (Ed.). PLoS ONE 8: e68535. https://doi.org/10.1371/journal.pone.0068535

Henrie JR, Thomson BM, Yungfleisch AA, Kerr M, Leavitt SD (2022) Characterizing Crustose Lichen Communities—DNA Metabarcoding Reveals More than Meets the Eye. Diversity 14: 766. https://doi.org/10.3390/d14090766

Hodgetts J, Ostojá-Starzewski JC, Prior T, Lawson R, Hall J, Boonham N (2016) DNA barcoding for biosecurity: case studies from the UK plant protection program. Wilson J-J (Ed.). Genome 59: 1033–1048. https://doi.org/10.1139/gen-2016-0010

Javal M, Terblanche J, Conlong D, Delahaye N, Grobbelaar E, Benoit L, Lopez-Vaamonde C, Haran J (2021) DNA barcoding for bio-surveillance of emerging pests and species identification

in Afrotropical Prioninae (Coleoptera, Cerambycidae). Biodiversity Data Journal 9: e64499. https://doi.org/10.3897/BDJ.9.e64499

Kelnarova I, Jendek E, Grebennikov VV, Bocak L (2019) First molecular phylogeny of *Agrilus* (Coleoptera: Buprestidae), the largest genus on Earth, with DNA barcode database for forestry pest diagnostics. Bulletin of Entomological Research 109: 200–211. https://doi.org/10.1017/S0007485318000330

Kenis M, Auger-Rozenberg M-A, Roques A, Timms L, Péré C, Cock MJW, Settele J, Augustin S, Lopez-Vaamonde C (2009) Ecological effects of invasive alien insects. Biological Invasions 11: 21–45. https://doi.org/10.1007/s10530-008-9318-y

Kitson JJN, Hahn C, Sands RJ, Straw NA, Evans DM, Lunt DH (2019) Detecting host–parasitoid interactions in an invasive Lepidopteran using nested tagging DNA metabarcoding. Molecular Ecology 28: 471–483. https://doi.org/10.1111/mec.14518

Koutroumpa FA, Lieutier F, Roux-Morabito G (2009) Incorporation of mitochondrial fragments in the nuclear genome (*Numts*) of the longhorned beetle *Monochamus galloprovincialis* (Coleoptera, Cerambycidae). Journal of Zoological Systematics and Evolutionary Research 47: 141–148. https://doi.org/10.1111/j.1439-0469.2008.00492.x

Krehenwinkel H, Pomerantz A, Prost S (2019) Genetic Biomonitoring and Biodiversity Assessment Using Portable Sequencing Technologies: Current Uses and Future Directions. Genes 10: 858. https://doi.org/10.3390/genes10110858

Liu M, Clarke LJ, Baker SC, Jordan GJ, Burridge CP (2020) A practical guide to DNA metabarcoding for entomological ecologists. Ecological Entomology 45: 373–385. https://doi.org/10.1111/een.12831

693 Mahé F, Rognes T, Quince C, de Vargas C, Dunthorn M (2014) Swarm: robust and fast clustering
694 method for amplicon-based studies. PeerJ 2: e593. https://doi.org/10.7717/peerj.593
695
696 Martin M (2011) Cutadapt removes adapter sequences from high-throughput sequencing
697 reads. EMBnet.journal 17: 10. https://doi.org/10.14806/ej.17.1.200
698
699 Mas H, Santoiemma G, Lencina JL, Gallego D, Pérez-Laorga E, Ruzzier E, Rassati D (2023)
700 Investigating beetle communities in and around entry points can improve surveillance at
701 national and international scale. NeoBiota 85: 145–165.
702 https://doi.org/10.3897/neobiota.85.103904
703
704 Mazza G, Tricarico E, Genovesi P, Gherardi F (2014) Biological invaders are threats to human
705 health: an overview. Ethology Ecology & Evolution 26: 112–129.
706 https://doi.org/10.1080/03949370.2013.863225
707
708 Nakahama N, Uchida K, Ushimaru A, Isagi Y (2018) Historical changes in grassland area
709 determined the demography of semi-natural grassland butterflies in Japan. Heredity 121: 155–
710 168. https://doi.org/10.1038/s41437-018-0057-2
711
712 Petec G, Vîlcu R, Uluitu M, Petec M (1975) Relationship between bovine serum albumin
713 structure and its chemical equilibria with hydrogen and 5-hydroxytryptamine ions. Physiologie
714 (Bucarest) 12: 77–80.
715
716 Piper AM, Batovska J, Cogan NOI, Weiss J, Cunningham JP, Rodoni BC, Blacket MJ (2019)
717 Prospects and challenges of implementing DNA metabarcoding for high-throughput insect
718 surveillance. GigaScience 8: giz092. https://doi.org/10.1093/gigascience/giz092
719
720 Ratnasingham S, Hebert PDN (2007) BARCODING: bold: The Barcode of Life Data System
721 (http://www.barcodinglife.org): BARCODING. Molecular Ecology Notes 7: 355–364.
722 https://doi.org/10.1111/j.1471-8286.2007.01678.x

Preston, M., Fritzsche, M. & Woodcock, P. 2022. Understanding and mitigating errors and biases in metabarcoding: an introduction for non-specialists. JNCC Report No. 699, JNCC, Peterborough, ISSN 0963-8091.

Richardson DM, Pysek P, Rejmanek M, Barbour MG, Panetta FD, West CJ (2000) Naturalization and invasion of alien plants: concepts and definitions. Diversity <html_ent glyph="@amp;" ascii="&"/> Distributions 6: 93–107. https://doi.org/10.1046/j.1472-4642.2000.00083.x

Roques A, Auger-Rozenberg M-A, Blackburn TM, Garnas J, Pyšek P, Rabitsch W, Richardson DM, Wingfield MJ, Liebhold AM, Duncan RP (2016) Temporal and interspecific variation in rates of spread for insect species invading Europe during the last 200 years. Biological Invasions 18: 907–920. https://doi.org/10.1007/s10530-016-1080-y

Roques A, Ren L, Rassati D, Shi J, Akulov E, Audsley N, Auger-Rozenberg M-A, Avtzis D, Battisti A, Bellanger R, Bernard A, Bernadinelli I, Branco M, Cavaletto G, Cocquempot C, Contarini M, Courtial B, Courtin C, Denux O, Dvořák M, Fan J, Feddern N, Francese J, Franzen EKL, Garcia A, Georgiev G, Georgieva M, Giarruzzo F, Gossner M, Gross L, Guarneri D, Hoch G, Hölling D, Jonsell M, Kirichenko N, Loomans A, Luo Y, McCullough D, Maddox C, Magnoux E, Marchioro M, Martinek P, Mas H, Mériguet B, Pan Y, Phélut R, Pineau P, Ray AM, Roques O, Ruiz M-C, Monteys VS i, Speranza S, Sun J, Sweeney JD, Touroult J, Valladares L, Veillat L, Yuan Y, Zalucki MP, Zou Y, Žunič-Kosi A, Hanks LM, Millar JG (2023) Worldwide tests of generic attractants, a promising tool for early detection of non-native cerambycid species. NeoBiota 84: 169–209. https://doi.org/10.3897/neobiota.84.91096

Rossa R, Goczał J (2021) Global diversity and distribution of longhorn beetles (Coleoptera: Cerambycidae). The European Zoological Journal 88: 289–302. https://doi.org/10.1080/24750263.2021.1883129

752 Sahlin K, Medvedev P (2020) De Novo Clustering of Long-Read Transcriptome Data Using a
753 Greedy, Quality Value-Based Algorithm. Journal of Computational Biology: A Journal of
754 Computational Molecular Cell Biology 27: 472–484. https://doi.org/10.1089/cmb.2019.0299
755
756 Seebens H, Bacher S, Blackburn TM, Capinha C, Dawson W, Dullinger S, Genovesi P, Hulme PE,
757 Kleunen M, Kühn I, Jeschke JM, Lenzner B, Liebhold AM, Pattison Z, Pergl J, Pyšek P, Winter M,
758 Essl F (2021) Projecting the continental accumulation of alien species through to 2050. Global
759 Change Biology 27: 970–982. https://doi.org/10.1111/gcb.15333
760
761 Seebens H, Blackburn TM, Dyer EE, Genovesi P, Hulme PE, Jeschke JM, Pagad S, Pyšek P, van
762 Kleunen M, Winter M, Ansong M, Arianoutsou M, Bacher S, Blasius B, Brockerhoff EG, Brundu
763 G, Capinha C, Causton CE, Celesti-Grapow L, Dawson W, Dullinger S, Economo EP, Fuentes N,
764 Guénard B, Jäger H, Kartesz J, Kenis M, Kühn I, Lenzner B, Liebhold AM, Mosena A, Moser D,
765 Nentwig W, Nishino M, Pearman D, Pergl J, Rabitsch W, Rojas-Sandoval J, Roques A, Rorke S,
766 Rossinelli S, Roy HE, Scalera R, Schindler S, Štajerová K, Tokarska-Guzik B, Walker K, Ward DF,
767 Yamanaka T, Essl F (2018) Global rise in emerging alien species results from increased
768 accessibility of new source pools. Proceedings of the National Academy of Sciences 115.
769 https://doi.org/10.1073/pnas.1719429115
770
771 Srivathsan A, Lee L, Katoh K, Hartop E, Kutty SN, Wong J, Yeo D, Meier R (2021) ONTbarcoder
772 and MinION barcodes aid biodiversity discovery and identification by everyone, for everyone.
773 BMC Biology 19: 217. https://doi.org/10.1186/s12915-021-01141-x
774
775 Strickler KM, Fremier AK, Goldberg CS (2015) Quantifying effects of UV-B, temperature, and
776 pH on eDNA degradation in aquatic microcosms. Biological Conservation 183: 85–92.
777 https://doi.org/10.1016/j.biocon.2014.11.038
778
779 Vamos E, Elbrecht V, Leese F (2017) Short COI markers for freshwater macroinvertebrate
780 metabarcoding. Metabarcoding and Metagenomics 1: e14625.
781 https://doi.org/10.3897/mbmg.1.14625

Vaser R, Sović I, Nagarajan N, Šikić M (2017) Fast and accurate de novo genome assembly from long uncorrected reads. Genome Research 27: 737–746. https://doi.org/10.1101/gr.214270.116

Wandeler P, Hoeck PEA, Keller LF (2007) Back to the future: museum specimens in population genetics. Trends in Ecology & Evolution 22: 634–642. https://doi.org/10.1016/j.tree.2007.08.017

Wang Q (2017) Cerambycidae of the World: Biology and Pest Management. CRC Press, 643 pp.

Wu Y, Trepanowski NF, Molongoski JJ, Reagel PF, Lingafelter SW, Nadel H, Myers SW, Ray AM (2017) Identification of wood-boring beetles (Cerambycidae and Buprestidae) intercepted in trade-associated solid wood packaging material using DNA barcoding and morphology. Scientific Reports 7: 40316. https://doi.org/10.1038/srep40316

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812 **Data accessibility**

813 Barcode data for the 32 species used in the mock community experiment are available from
814 BOLD in the dataset DS-MINION (dx.doi.org/XX/DS-MINION). Raw sequence data for this
815 project and analytical script and files are available on figshare
816 (https://figshare.com/projects/Molecular_biosurveillance_of_wood-
817 boring_cerambycid_beetles_using_DNA_metabarcoding/171432).

818

819 **Author contributions**

820 Loïs Veillat, Géraldine Roux, Carlos Lopez-Vaamonde and Stéphane Boyer conceived the study.
821 Alain Roques collected field samples. Stéphane Boyer, Marina Querejeta, Emmanuelle
822 Magnoux and Loïs Veillat conducted the laboratory sample processing. Loïs Veillat analysed
823 the data and wrote the first draft. All authors contributed to the preparation of the manuscript.
824 Both senior authors, Géraldine Roux and Carlos Lopez-Vaamonde, contributed equally to this
825 study.

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841 Figure 1: Overview of the DNA extraction protocol for tests 1 and 3 (Fig. 1a) and for test 2 (Fig.
842 1b).

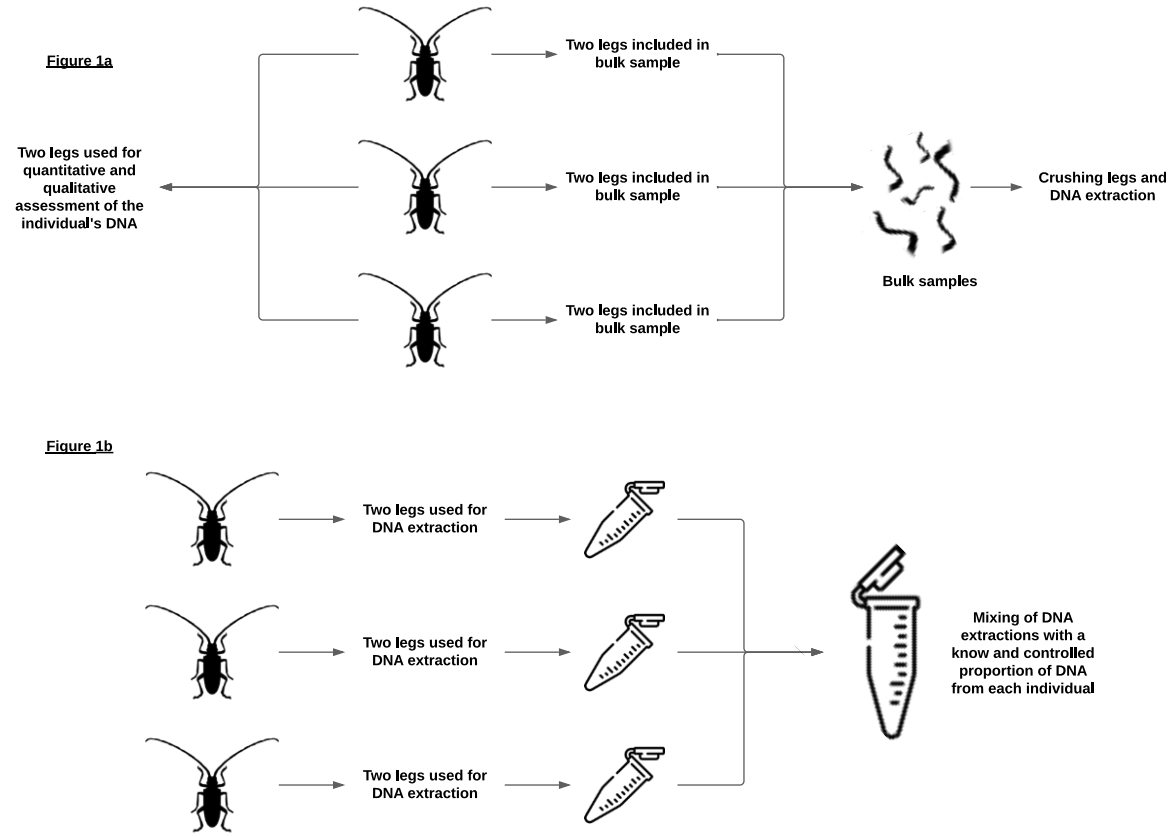

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Figure 2: Venn diagrams showing the number of specific and shared species-level detections among primers used (upper, F=fwhF2/fwhR2n [254bp] and B=BF3/BR2 [458bp]), technology used (middle) and for primers and technologies used (lower).

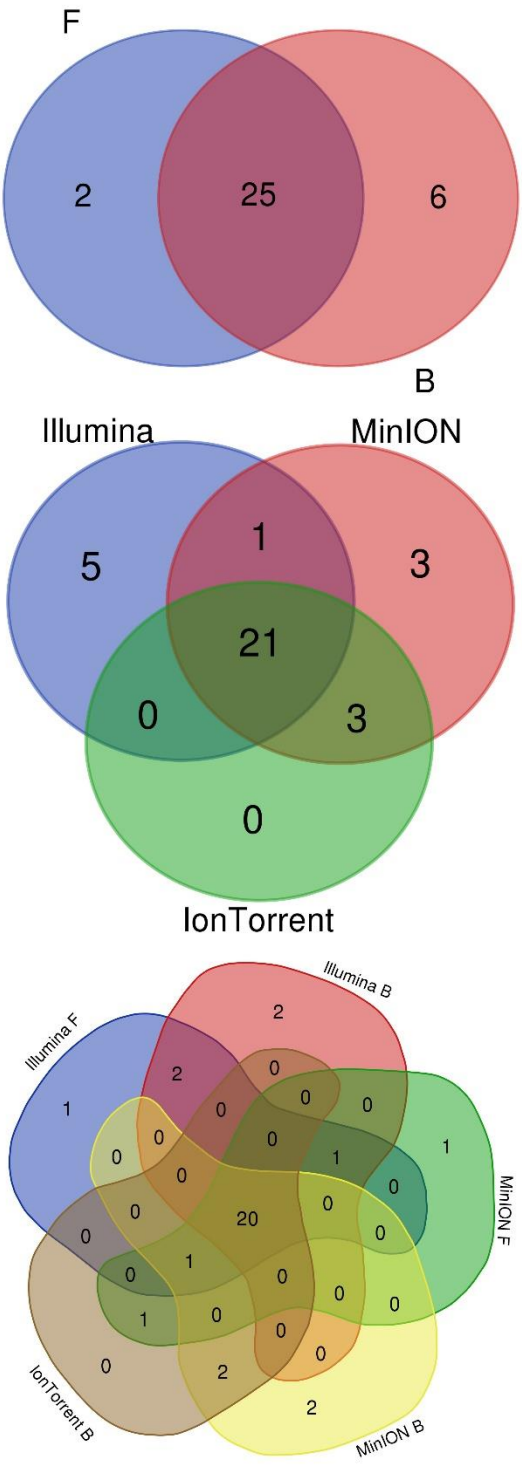
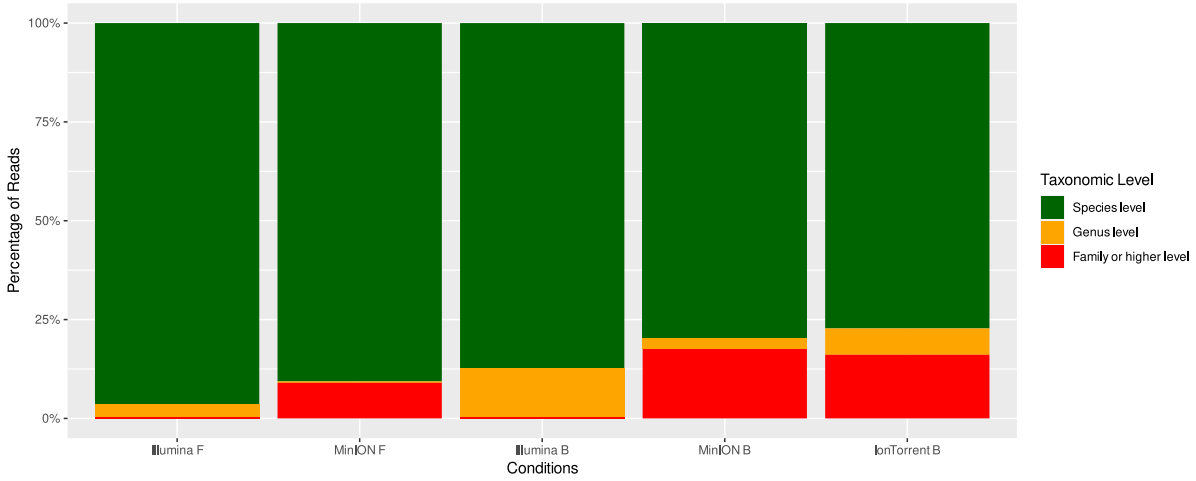
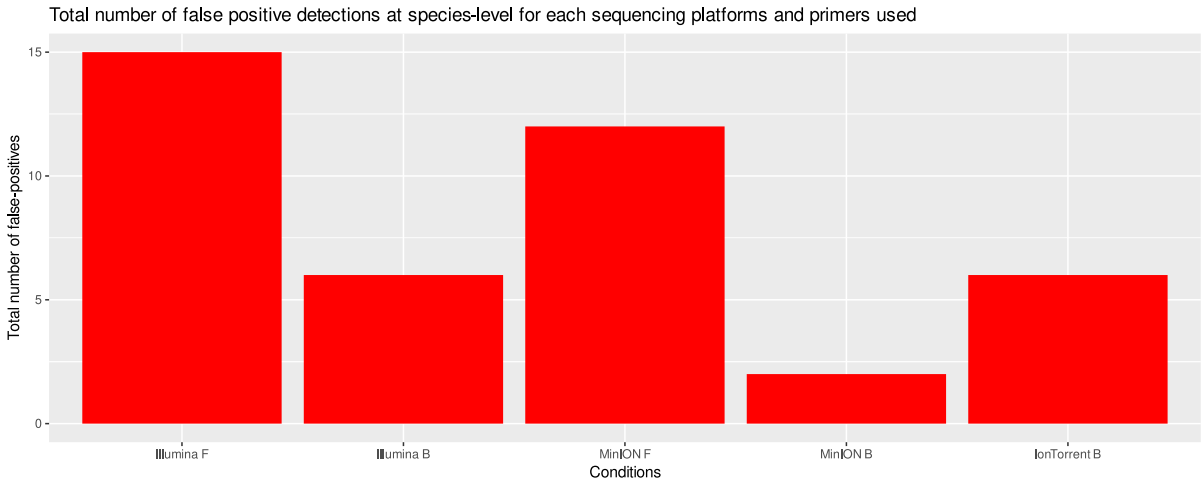


Figure 3: Proportion of reads assigned to each taxonomic level for each combination of sequencing technology and pair of primers (F: fwhF2/fwhR2n; B: BF3/BR2).



884 Figure 4: Number of false positive detections at species-level for each sequencing platforms
885 and primers used (F=fwhF2/fwhR2n [254bp] and B=BF3/BR2 [458bp]).
886



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907 Figure 5: Heatmap comparing the identification of individuals present in bulk samples at the
908 species level (green square) or the absence of detection at the species level (red square)
909 according to the sequencing technologies and primer pairs used. Species names written in blue
910 were collected using the wet method, those in yellow were collected using the dry method,
911 and those in dark green were hand-captured.
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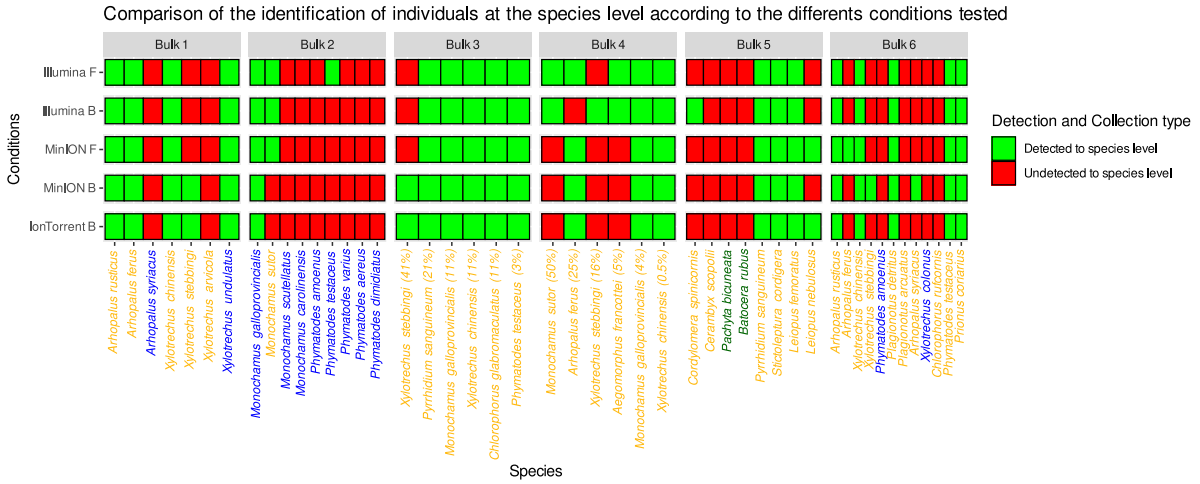

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Figure 6: Boxplots representing (A) the average number of detections according to the type of preservation used, (B) the logarithm of the average DNA concentration according to the type of preservation used, (C) the A260/280 quality ratio according to the type of preservation used, and (D) the A260/230 quality ratio according to the type of preservation used. The black dots represent the extreme values.