

PREPRINT

Author-formatted, not peer-reviewed document posted on 21/03/2024

DOI: <https://doi.org/10.3897/arphapreprints.e123465>

Resolving the issues of translocated native species in freshwater invasions

 Ali Serhan Tarkan,  Irmak Kurtul, Dagmara Blonska,  John Robert Britton,  Phillip Haubrock

Resolving the issues of translocated native species in freshwater invasions

Ali Serhan Tarkan^{1,2,3*}, Irmak Kurtul^{2,4}, Dagmara Błońska^{1,2}, J. Robert Britton², Phillip J. Haubrock^{5,6,7}

¹Department of Ecology and Vertebrate Zoology, Faculty of Biology and Environmental Protection, University of Lodz, Lodz, Poland

²Department of Life and Environmental Sciences, Faculty of Science and Technology, Bournemouth University, Poole, Dorset, The United Kingdom

³Department of Basic Sciences, Faculty of Fisheries, Muğla Sıtkı Koçman University, Muğla, Türkiye

⁴Marine and Inland Waters Sciences and Technology Department, Faculty of Fisheries, Ege University, İzmir, Türkiye

⁵Senckenberg Research Institute and Natural History Museum Frankfurt, Department of River Ecology and Conservation, Gelnhausen, Germany

⁶University of South Bohemia in České Budějovice, Faculty of Fisheries and Protection of Waters, South Bohemian Research Centre of Aquaculture and Biodiversity of Hydrocenoses, Zátiší 728/II, 389 25 Vodňany, Czech Republic

⁷Center for Applied Mathematics and Bioinformatics, Department of Mathematics and Natural Sciences, Gulf University for Science and Technology, Hawally, Kuwait

*Corresponding author: Ali Serhan Tarkan; serhantarkan@gmail.com

Abstract

Biological invasions, driven by human-mediated species movements, pose significant threats to global ecosystems and economies. The classification of non-native species is a complex issue intertwining ecological considerations and ethical concerns. The need for nuanced and less ambiguous terminology is emphasised, considering biogeographic, evolutionary, and ecological principles. In-country translocations of native species into ecosystems they do not naturally occur, are often overlooked and are the least regulated among species movements, despite being increasingly common in conservation. Our case studies, spanning various ecosystems and taxa, illustrate the diverse impacts of translocations on native species and ecosystems. The challenges associated with translocated species underscore the urgency for robust risk management strategies and rigorous monitoring. A comprehensive and adaptable management framework that considers translocated species for evidence-based management decisions is critical for navigating the complexities of translocations effectively, ensuring the conservation of biodiversity and ecosystem sustainability.

Keywords: biological invasions, conservation, translocations, invasive species management

Introduction

Biological invasions are where species are moved by human activities from their native range to new areas where they have no evolutionary history and are a major global economic and ecological concern (Simberloff et al. 2005). Biological invasions are recognized as a pervasive threat to biodiversity and human well-being, especially in aquatic ecosystems (Cuthbert et al. 2021; Haubrock et al. 2021). Numerous pathways facilitate the spread of aquatic non-native species (Ruiz et al. 2011), with human-mediated pathways involving global trade and transportation (Avila et al. 2020). Once established in new habitats, aquatic non-native species can disrupt local ecosystems by competing with native species for resources, modifying habitats, and altering nutrient cycles, often with severe consequences for human well-being (Piria et al. 2021; Hald-Mortensen 2023). Economically, these species can damage sectors including fisheries and, among others, public welfare, leading to substantial financial costs (Cuthbert et al. 2021; Warziniack et al. 2021).

Terminology of non-native species

The classification of non-native species is a complex issue intertwined with both ecological considerations and ethical concerns (Richardson et al. 2011) and was recently reviewed and discussed by Soto et al. (2024). Ethically, the language used in invasion science to describe non-native species often mirrors societal views on foreignness, with terms for non-native species such as “alien” reflecting and even reinforcing xenophobic attitudes (Soto et al. 2024), paralleling language used against human immigrants. From an ecological perspective, the current classification systems used for ‘invasiveness’ can differ substantially, where the focus is the ecological impact of the species or its ability to establish and spread, but rarely both (Soto et al. 2024).

In general, policies and management of non-native species rely strongly on national boundaries (Piria et al. 2021), which can be highly problematic as these boundaries fail to consider biogeographic principles on the evolutionary history and intricate ecological interactions of the species being moved (Soto et al. 2024). This is especially true for freshwater ecosystems, where nativeness and non-nativeness may even differ between adjacent river basins within the same biogeographic realm (e.g Warren et al. 2024). This is then exacerbated by ambiguous terminology with, for example, the EU Invasive Alien Species Regulation (Regulation (EU) 1143/2014) definition of an “invasive alien species” being ‘*animals and plants that are introduced accidentally or deliberately into a natural environment where they are not normally found, with serious negative consequences for their new environment*’, with the aim of the associated legislation to ‘*prevent, minimise and mitigate the adverse impacts posed by these species on native biodiversity and ecosystem services*’. Accordingly, there remains considerable ambiguity as to what constitutes the ‘*natural environment*’ and ‘*native biodiversity*’ (Chew and Hamilton 2011). We thus argue the requirement for a more nuanced and less ambiguous terminology for native species translocated within their native region that combines the issues of the species’ biogeographic region of origin and ecosystems it naturally occurs in, as well as its evolutionary history and ecological role(s), rather than just relying on geo-political (often national) boundaries. More explicitly, we suggest that current terminology, such as “native biodiversity” and “non-native species” when considered in the assessment of biological invasions is too associated with national boundaries and their use must instead be based on sound biogeographical, evolutionary and ecological principles (Nehring and Klingenstein 2008; Wolter and Röhr 2010).

The abuse of terminology in this regard is well illustrated by native species that can become pests in their native region (previously named “native invasive”, Simberloff and Rejmánek 2011; see also Soto et al. 2024) through human activities. This can occur through

anthropogenic driven environmental changes that result in the abundance of native species rapidly increasing through shifts in life history traits, and through human-induced habitat changes that create novel environments where some native species can form highly abundant populations to the detriment of others (e.g. the transformation of lotic to lentic environments through impoundment; Šmejkal et al. 2023). It can also occur as a direct result of human actions, including intentional stocking, where a species considered native (according to national boundaries) is moved into a new area within either their original biogeographical range or into a new biogeographic range within the country (Carey et al. 2012). We argue this activity, despite often practised by conservationists to e.g. protect a highly endangered species (Ricciardi and Simberloff 2009; Olden et al. 2011; Bradley et al. 2022), represents the release of a non-native species and potentially results in a biological invasion whose harm on the receiving ecosystem is likely to be underestimated (Gilroy et al. 2017). It is these releases of species within national boundaries, but between biogeographic areas that ignore evolutionary histories and ecological roles, which we consider as being highly problematic for invasion management (Usher 2000; 2020; Soto et al. 2024).

The issue of in-country translocations: case studies from freshwater ecosystems

Efforts to conserve biodiversity and the aim of invasion scientists to understand and mitigate biological invasions are often perceived as a philosophical paradox due to synergistic overlaps concomitant to differing priorities (i.e. species native in one region but invasive in another; Pérez et al. 2006; Marchetti and Engstrom 2016). In recent years, the importance of in-country translocations of native species have increased in conservation worldwide (Vitule et al. 2019), yet are still largely overlooked by invasion scientists (see Glamuzina et al. 2017), despite being particularly common in certain countries (Tarkan et al. 2017). However, while conservation

related translocations are often pre-planned and strictly regulated, movements of species for use in fisheries, aquaculture and the ornamental trade are less regulated, leading to widespread secondary spread (Vander Zanden and Olden 2008). Translocated native and non-native species (*sensu stricto*) thus can pose a considerable threat to native species and ecosystems, especially where the translocation has been poorly regulated (e.g. Hodder and Bullock 1997; Glamuzina et al. 2017), which we demonstrate in the following case studies.

Translocated fishes

Translocations of freshwater fishes are commonplace, as this easily completed exercise can be used to enhance aquaculture production, and catches in commercial, artisanal and recreational fisheries. It has been used extensively in East Africa, with species such as Nile tilapia *Oreochromis niloticus* moved extensively between lakes in Kenya to enhance fish catches and improve food security (Geletu and Zhao 2023). These translocations have contributed to fish diversity loss in recent years, including through their hybridisation with native congeners, with the interaction of their translocation dynamics with aquaculture escapes also driving artificial gene flow between different Nile tilapia stocks, impacting the integrity of local gene pools (such as through outbreeding depression), impacting the sustainability of the species as a resource for fisheries (Tibihika et al. 2022).

In England, fish species richness is naturally higher in eastern flowing rivers than those flowing west. This resulted from a now drowned land-bridge with mainland Europe at the end of the last glacial period that connected these eastern flowing rivers with the Rhine and Danube systems, providing a route for fish recolonisation from glacial refuges further south (Wheeler 1977). In the last 100 years, there has been the frequent translocation of species, such as European barbel *Barbus barbus*, from these eastern flowing rivers where they are indigenous

(usually in the Thames basin in southeastern England) to the western flowing rivers, where they are non-indigenous (e.g. Wheeler and Jordan 1990). A prominent example of this was the translocation of 509 adult fish from the River Kennet into the middle reaches of the River Severn in 1956, and completed by the fishery regulator of that time with the aim to enhance angling (Wheeler and Jordan 1990; Antognazza et al. 2016). These fish rapidly established a sustainable population which dispersed throughout the Severn basin and also resulted in further translocations in western Britain, with anglers moving these fish to neighbouring basins, such as the River Wye (Antognazza et al. 2016). In addition, translocations in the indigenous range involve the movement of hatchery reared barbel reared using broodstock from one basin (often the Thames again) and releasing them in different basins, with this already identified as impacting barbel genetic integrity in northeast England (Antognazza et al. 2016). Accordingly, barbel are now widespread through Great Britain, with populations in England, Scotland and Wales due to translocations, despite their native range being restricted to a small number of basins in eastern England (Wheeler and Jordan 1990; Antognazza et al. 2016).

European perch (*Perca fluviatilis*), known for its aggressive feeding behaviour that often results in the extirpation of native fish species, has been extensively translocated between different bodies of water within its native range by anglers from Thrace (European part of Turkey) to newly established water reservoirs in the Anatolian part (Tarkan et al. 2023b). The translocated perch exhibit higher aggression levels than native populations, impacting native fish communities (Tarkan et al. 2023a), and potentially lead to cascading effects throughout the food web, altering community structure and ecosystem dynamics, with implications for both ecosystem functioning and human well-being (Tarkan et al. 2023a). Similarly, the extirpation of two endemic fish species in lakes Egirdir and Beysehir (southern Anatolia) has been linked to the introduction of translocated piscivorous pikeperch *Sander lucioperca* (Tarkan et al. 2014).

A unique example in support of our argument relates to the existence of two distinctive populations of racer goby (*Babka gymnotrachelus*) in Poland. In the mid-1990s, the species was recorded in the Vistula drainage system, likely reaching it from the Dnieper through Pripyat-Bug canals (Semenchenko et al. 2011). It has since been listed among other spreading non-native species in Polish rivers (Grabowska et al. 2010). However, monitoring studies in 2009 in the Strwiąż River, a tributary of the upper Dniester River, identified an abundant population of racer goby, suggesting its native status in Poland. As genetic analyses confirmed the dual origin of the species (Grabowski et al. 2015), this creates an ambiguous situation where, considering administrative borders, the species is simultaneously native to one and invasive to another tributary within the same country, posing challenges from a legislative and regulatory perspective.

Other translocated taxa

The issue of translocations is not limited to fish but is a cross-taxa issue involving amphibians, reptiles and crustaceans. For crustaceans, the translocation of the freshwater shrimp (*Paratya australiensis*) within the same drainage system in Australia to maintain and even increase genetic diversity led to the extirpation of the resident genotype within seven years due to mating preferences of females with translocated males and the low viability of crosses between resident females and translocated males (Hughes et al. 2003). In Australia, the translocation of three native freshwater crayfish species (*Cherax tenuimanus*, *C. destructor*, and *C. quadricarinatus*) raised concerns due to the subsequent harmful impacts on native freshwater ecosystems (Beatty et al. 2005). While Australia has established controls to manage the import and export of these crayfish, the regulatory approach within the country lacked uniformity and, ultimately, led to numerous impacts, including the introduction of diseases, disruption of local

ecosystems through competitive interactions with native species, habitat alterations, and genetic dilution through hybridisation. Similar cases may also be found in North America, where many crayfish species are widespread, but where the native regions and river basins do not overlap with state boundaries (Taylor et al. 2007).

For other taxa, translocations often have negative outcomes for the released individuals rather than resulting in invasions, which can be problematic if the driver of the translocation was to relocate endangered animals (such as amphibians) that are under threat from habitat destruction (Bradley et al. 2022). Such translocations for mitigation effects are a form of assisted colonisation and mirror debates on using this as a climate change adaptation action for protecting vulnerable species (e.g. Lunt et al. 2013). To reduce human-wildlife conflicts reptiles are moved to new locations where they seem to experience elevated mortality rates compared to resident individuals. This is frequently linked to unusual movement patterns, stress, disease, and challenges in surviving winters, particularly for species that prioritise locating suitable hibernation sites (summarised by Sullivan et al. 2015; Cornelis et al. 2021).

Redefining 'native area': a call for a biogeographic ecosystem approach

These case studies indicate that the translocation of species between river basins may exhibit diverse reactions based on the specific environmental conditions in which they are introduced (Tarkan et al. 2017). This inherent variability emphasises the need for a nuanced understanding of ecological dependencies, as not all translocated species respond uniformly to non-native counterparts (Vitule et al. 2019). The underlying factor driving such varied responses lies in the ecological dependency of species, whereby their behaviour is intricately influenced by the environmental context (Strona et al. 2021).

Accordingly, we argue that the issue of translocated native species within national boundaries demands a re-evaluation of the concept of 'native area' (Guichón et al. 2015) and associated terminology (Soto et al. 2024), particularly in the contexts of fisheries, aquaculture, and the ornamental trade. Traditional classification systems based on national boundaries are insufficient for addressing the ecological complexities of species translocations (Pyšek et al. 2004). A bio-geographically informed approach, recognizing the ecological and evolutionary contexts of species, is imperative. These could, among others, include river basin district (RBD) type approaches as implemented in the Water Framework Directive (Nilsson et al. 2004) and thus, we emphasise that our primary concern lies with the movements for fisheries, aquaculture, and ornamental trade, areas where risk screening and regulatory measures are more strictly adopted. The implementation of such an approach would involve considering the historical distribution, ecological interactions, and evolutionary relationships of species to define their nativity more accurately. This shift in perspective would enable conservationists and policymakers to develop more effective and ecologically sound strategies for managing non-native species and allow a more accurate risk screening and assessment process (Copp et al. 2016; Tarkan et al. 2020). We then suggest the term 'translocated native species' is replaced by 'introduced native species', bringing it in-line with invasion science terminology and embedding biogeographic principles within policy and regulation.

Conclusion

The evident importance of species translocated within their native region in the context of biological invasions, equivalent to that of non-native species, highlights the need for a flexible management framework designed to fully incorporate and address the nuances of species propagated for commercial sale. Such a framework should consider both the native

species natural ecosystems, biogeographic distribution, and evolutionary history when outlining its natural occurrences. One such framework could be the Dispersal-Origin-Status-Impact (DOSI) assessment scheme, introduced by (Soto et al. 2024) DOSI classifies populations of non-native species at the population level. For this, it assesses non-native species based on their dispersal methods (assisted or independent), origin (allochthonous or autochthonous), current status (expanding, stationary, or shrinking), and impact (ecological, economic, health, or cultural). DOSI's flexible and comprehensive approach supports objective, data-driven decision-making for managing biological invasions, allowing for prioritisation of interventions at various scales. This method represents an improvement over previous strategies by addressing the needs of managers and stakeholders with limited resources. DOSI could be expanded to include introduced species (i.e., species translocated within their native range to ecosystems where they do not occur naturally) or native pests whose inclusion might refine the management strategies under DOSI. DOSI only considers negative impacts (i.e., potential threats), acknowledging that negative impacts considerably outweigh and are distinct from any potential benefits. However, the aim of DOSI is to prioritise populations of non-native species for management interventions based on local risks, disregarding the feasibility or existence of adequate approaches, and the species' ability to spread beyond current confinements. While this is one possibility, the intricate challenges associated with translocated species however stress the urgency for robust risk management strategies, complemented by meticulous monitoring and centralised databases, to navigate the complexities of translocations in more effective ways.

Acknowledgements

We dedicate this work to the memory of Prof Gordon H Copp who passed away on 8 July 2023 for his inspirational work and enthusiasm on invasion biology and risk management.

References

- Antognazza CM, Andreou D, Zaccara S, Britton RJ (2016) Loss of genetic integrity and biological invasions result from stocking and introductions of *Barbus barbus*: insights from rivers in England. *Ecology and Evolution* 6: 1280–1292. <https://doi.org/10.1002/ece3.1906>
- Avila C, Angulo-Preckler C, Martín-Martín RP, Figuerola B, Griffiths HJ, Waller CL (2020) Invasive marine species discovered on non-native kelp rafts in the warmest Antarctic island. *Scientific Reports* 10: 1639. <https://doi.org/10.1038/s41598-020-58561-y>
- Beatty S, Morgan D, Gill H (2005) Role of Life History Strategy in the Colonisation of Western Australian Aquatic Systems by the Introduced Crayfish *Cherax destructor* Clark, 1936. *Hydrobiologia* 549: 219–237. <https://doi.org/10.1007/s10750-005-5443-0>
- Bradley HS, Tomlinson S, Craig MD, Cross AT, Bateman PW (2022) Mitigation translocation as a management tool. *Conservation Biology* 36. <https://doi.org/10.1111/cobi.13667>
- Carey MP, Sanderson BL, Barnas KA, Olden JD (2012) Native invaders – challenges for science, management, policy, and society. *Frontiers in Ecology and the Environment* 10: 373–381. <https://doi.org/10.1890/110060>
- Chew MK, Hamilton AL (2011) The rise and fall of biotic nativeness: a historical perspective. *Fifty years of invasion ecology: the legacy of Charles Elton*: 35–48.
- Copp G, Vilizzi L, Tidbury H, Stebbing P, Tarkan AS, Miossec L, Gouilletquer P (2016) Development of a generic decision-support tool for identifying potentially invasive

283 aquatic taxa: AS-ISK. Management of Biological Invasions 7: 343–350.
 284 <https://doi.org/10.3391/mbi.2016.7.4.04>

285 Cornelis J, Parkin T, Bateman PW (2021) Killing them softly: a review on snake translocation
 286 and an Australian case study. Herpetological Journal: 118–131.
 287 <https://doi.org/10.33256/31.3.118131>

288 Cuthbert RN, Pattison Z, Taylor NG, Verbrugge L, Diagne C, Ahmed DA, Leroy B, Angulo
 289 E, Briski E, Capinha C (2021) Global economic costs of aquatic invasive alien species.
 290 Science of the Total Environment 775: 145238.

291 Geletu TT, Zhao J (2023) Genetic resources of Nile tilapia (*Oreochromis niloticus* Linnaeus,
 292 1758) in its native range and aquaculture. Hydrobiologia 850: 2425–2445.
 293 <https://doi.org/10.1007/s10750-022-04989-4>

294 Gilroy JJ, Avery JD, Lockwood JL (2017) Seeking International Agreement on What it Means
 295 To be “Native.” Conservation Letters 10: 238–247. <https://doi.org/10.1111/conl.12246>

296 Glamuzina B, Tutman P, Nikolić V, Vidović Z, Pavličević J, Vilizzi L, Copp GH, Simonović
 297 P (2017) Comparison of Taxon-Specific and Taxon-Generic Risk Screening Tools to
 298 Identify Potentially Invasive Non-native Fishes in the River Neretva Catchment (Bosnia
 299 and Herzegovina and Croatia). River Research and Applications 33: 670–679.
 300 <https://doi.org/10.1002/rra.3124>

301 Grabowska J, Kotusz J, Witkowski A (2010) Alien invasive fish species in Polish waters: an
 302 overview. Folia Zoologica 59: 73–85. <https://doi.org/10.25225/fozo.v59.i1.a1.2010>

303 Grabowski M, Hupał K, Bylak A, Kukuła K, Grabowska J (2015) Double origin of the racer
 304 goby (*Babka gymnotrachelus*) in Poland revealed with mitochondrial marker. Possible
 305 implications for the species alien/native status. Journal of Limnology.
 306 <https://doi.org/10.4081/jlimnol.2015.1253>

- 307 Guichón ML, Benitez V V., Gozzi AC, Hertzriken M, Borgnia M (2015) From a lag in vector
308 activity to a constant increase of translocations: invasion of *Callosciurus squirrels* in
309 Argentina. *Biological Invasions* 17: 2597–2604. [https://doi.org/10.1007/s10530-015-](https://doi.org/10.1007/s10530-015-0897-0)
310 0897-0
- 311 Hald-Mortensen C (2023) The Main Drivers of Biodiversity Loss: A Brief Overview. *Journal*
312 *of Ecology and Natural Resources* 7: 000346.
- 313 Haubrock PJ, Pilotto F, Innocenti G, Cianfanelli S, Haase P (2021) Two centuries for an almost
314 complete community turnover from native to non-native species in a riverine ecosystem.
315 *Global Change Biology* 27: 606–623.
- 316 Hodder KH, Bullock JM (1997) Translocations of Native Species in the UK: Implications for
317 Biodiversity. *The Journal of Applied Ecology* 34: 547. <https://doi.org/10.2307/2404906>
- 318 Hughes J, Goudkamp K, Hurwood D, Hancock M, Bunn S (2003) Translocation Causes
319 Extinction of a Local Population of the Freshwater Shrimp *Paratya australiensis*.
320 *Conservation Biology* 17: 1007–1012. <https://doi.org/10.1046/j.1523-1739.2003.01636.x>
- 321 Lunt ID, Byrne M, Hellmann JJ, Mitchell NJ, Garnett ST, Hayward MW, Martin TG,
322 McDonald-Maddden E, Williams SE, Zander KK (2013) Using assisted colonisation to
323 conserve biodiversity and restore ecosystem function under climate change. *Biological*
324 *Conservation* 157: 172–177. <https://doi.org/10.1016/j.biocon.2012.08.034>
- 325 Marchetti MP, Engstrom T (2016) The conservation paradox of endangered and invasive
326 species. *Conservation Biology* 30: 434–437.
- 327 Nehring S, Klingenstein F (2008) Aquatic alien species in Germany—listing system and options
328 for action. *Neobiota* 7: 19–33.
- 329 Nilsson S, Langaas S, Hannerz F (2004) International river basin districts under the EU Water
330 Framework Directive: Identification and planned cooperation. *European Water*
331 *Management Online* 2: 1–20.

- 332 Olden JD, Kennard MJ, Lawer JJ, Poff NL (2011) Challenges and Opportunities in
 333 Implementing Managed Relocation for Conservation of Freshwater Species. *Conservation*
 334 *Biology* 25: 40–47. <https://doi.org/10.1111/j.1523-1739.2010.01557.x>
- 335 Pérez JE, Nirchio M, Alfonsi C, Muñoz C (2006) The biology of invasions: the genetic
 336 adaptation paradox. *Biological Invasions* 8: 1115–1121.
- 337 Piria M, Stroil BK, Giannetto D, Tarkan AS, Gavrilović A, Špelić I, Radočaj T, Killi N, Filiz
 338 H, Uysal TU, Aldemir C, Kamberi E, Hala E, Bakiu R, Kolutari J, Buda E, Bakiu SD,
 339 Sadiku E, Bakrač A, Mujić E, Avdić S, Doumpas N, Giovos I, Dinoshi I, Ušanović L,
 340 Kalajdžić A, Pešić A, Četković I, Marković O, Milošević D, Mrdak D, Sará G, Belmar
 341 MB, Marchessaux G, Trajanovski S, Zdraveski K (2021) An assessment of regulation,
 342 education practices and socio-economic perceptions of non-native aquatic species in the
 343 Balkans. *Journal of Vertebrate Biology* 70. <https://doi.org/10.25225/jvb.21047>
- 344 Pyšek P, Richardson DM, Rejmánek M, Webster GL, Williamson M, Kirschner J (2004) Alien
 345 plants in checklists and floras: towards better communication between taxonomists and
 346 ecologists. *TAXON* 53: 131–143. <https://doi.org/10.2307/4135498>
- 347 Ricciardi A, Simberloff D (2009) Assisted colonization is not a viable conservation strategy.
 348 *Trends in Ecology & Evolution* 24: 248–253. <https://doi.org/10.1016/j.tree.2008.12.006>
- 349 Richardson DM, Pyšek P, Carlton JT (2011) A compendium of essential concepts and
 350 terminology in invasion ecology. Fifty years of invasion ecology: the legacy of Charles
 351 Elton 1: 409–420.
- 352 Ruiz GM, Fofonoff PW, Steves B, Foss SF, Shiba SN (2011) Marine invasion history and
 353 vector analysis of California: a hotspot for western North America. *Diversity and*
 354 *Distributions* 17: 362–373. <https://doi.org/10.1111/j.1472-4642.2011.00742.x>

- 355 Semenchenko V, Grabowska J, Grabowski M, Rizevsky V, Pluta M (2011) Non-native fish in
356 Belarusian and Polish areas of the European central invasion corridor. *Oceanological and*
357 *Hydrobiological Studies* 40: 57–67. <https://doi.org/10.2478/s13545-011-0007-6>
- 358 Simberloff D, Rejmánek M (2011) *Encyclopedia of biological invasions*. Univ of California
359 Press.
- 360 Simberloff D, Parker IM, Windle PN (2005) Introduced species policy, management, and
361 future research needs. *Frontiers in Ecology and the Environment* 3: 12–20.
- 362 Šmejkal M, Bartoň D, Duras J, Horký P, Muška M, Kubečka J, Pfauserová N, Tesfaye M,
363 Slavík O (2023) Living on the edge: Reservoirs facilitate enhanced interactions among
364 generalist and rheophilic fish species in tributaries. *Frontiers in Environmental Science*
365 11. <https://doi.org/10.3389/fenvs.2023.1099030>
- 366 Soto I, Balzani P, Carneiro L, Cuthbert RN, Macedo R, Tarkan AS, Ahmed DA, Bang A,
367 Bacela-Spychalska K, Bailey SA (2024) Taming the terminological tempest in invasion
368 science. <https://doi.org/10.1002/BRV.13071>
- 369 Strona G, Beck PSA, Cabeza M, Fattorini S, Guilhaumon F, Micheli F, Montano S, Ovaskainen
370 O, Planes S, Veech JA, Parravicini V (2021) Ecological dependencies make remote reef
371 fish communities most vulnerable to coral loss. *Nature Communications* 12: 7282.
372 <https://doi.org/10.1038/s41467-021-27440-z>
- 373 Sullivan BK, Nowak EM, Kwiatkowski MA (2015) Problems with mitigation translocation of
374 herpetofauna. *Conservation Biology* 29: 12–18. <https://doi.org/10.1111/cobi.12336>
- 375 Tarkan AS, Güler Ekmekçi F, Vilizzi L, Copp GH (2014) Risk screening of non-native
376 freshwater fishes at the frontier between Asia and Europe: first application in Turkey of
377 the fish invasiveness screening kit. *Journal of Applied Ichthyology* 30: 392–398.
378 <https://doi.org/10.1111/jai.12389>

- 379 Tarkan AS, Vilizzi L, Top N, Ekmekçi FG, Stebbing PD, Copp GH (2017) Identification of
380 potentially invasive freshwater fishes, including translocated species, in Turkey using the
381 Aquatic Species Invasiveness Screening Kit (AS-ISK). International Review of
382 Hydrobiology 102: 47–56. <https://doi.org/10.1002/iroh.201601877>
- 383 Tarkan AS, Yoğurtçuoğlu B, Ekmekçi FG, Clarke SA, Wood LE, Vilizzi L, Copp G (2020)
384 First application in Turkey of the European Non-native Species in Aquaculture Risk
385 Analysis Scheme to evaluate the farmed non-native fish, striped catfish *Pangasianodon*
386 *hypophthalmus*. Fisheries Management and Ecology 27: 123–131.
387 <https://doi.org/10.1111/fme.12387>
- 388 Tarkan AS, Haubrock PJ, Aksu S, Mol O, Balzani P, Emiroğlu Ö, Köse E, Kurtul I, Başkurt S,
389 Çınar E, Oztopcu-Vatan P (2023a) Predicting the potential implications of perch (*Perca*
390 *fluviatilis*) introductions to a biodiversity-rich lake using stable isotope analysis. Scientific
391 Reports 13: 17635. <https://doi.org/10.1038/s41598-023-44865-2>
- 392 Tarkan AS, Mol O, Aksu S, Köse E, Kurtul I, Başkurt S, Haubrock PJ, Balzani P, Çınar E,
393 Britton JR, Oztopcu-Vatan P, Emiroğlu Ö (2023b) Phenotypic responses to piscivory in
394 invasive gibel carp populations. Aquatic Sciences 85: 75. [https://doi.org/10.1007/s00027-](https://doi.org/10.1007/s00027-023-00974-8)
395 [023-00974-8](https://doi.org/10.1007/s00027-023-00974-8)
- 396 Taylor CA, Schuster GA, Cooper JE, DiStefano RJ, Eversole AG, Hamr P, Hobbs III HH,
397 Robison HW, Skelton CE, Thoma RF (2007) A reassessment of the conservation status
398 of crayfishes of the United States and Canada after 10+ years of increased awareness.
399 Fisheries 32: 372–389.
- 400 Tibihika PD, Meimberg H, Curto M (2022) Understanding the translocation dynamics of Nile
401 tilapia (*Oreochromis niloticus*) and its ecological consequences in East Africa. African
402 Zoology 57: 171–179. <https://doi.org/10.1080/15627020.2022.2154169>
- 403 Usher M (2020) Territory incognita. Progress in Human Geography 44: 1019–1046.

- Usher MB (2000) The nativeness and non-nativeness of species. *Watsonia* 23: 323–326.
- Vitule JRS, Occhi TVT, Kang B, Matsuzaki S-I, Bezerra LA, Daga VS, Faria L, Frehse F de A, Walter F, Padial AA (2019) Intra-country introductions unraveling global hotspots of alien fish species. *Biodiversity and Conservation* 28: 3037–3043. <https://doi.org/10.1007/s10531-019-01815-7>
- Warren BIC, Pinder AC, Parker B, Tarkan AS, Britton JR (2024) Trophic relationships of translocated and indigenous chub *Squalius cephalus* populations with trophically analogous fishes. *Hydrobiologia* 851: 1291–1303. <https://doi.org/10.1007/s10750-023-05389-y>
- Warziniack T., Haight R. G., Yemshanov D., Apriesnig J. L., Holmes T. P., Countryman A. M., Haberland C. (2021) Economics of invasive species. In: Poland T. M. P-WT, FDM, MCF, HDC, LVM (Ed.), *Invasive species in forests and rangelands of the United States: a comprehensive science synthesis for the United States forest sector*. Springer, 305–320.
- Wheeler A (1977) The Origin and Distribution of the Freshwater Fishes of the British Isles. *Journal of Biogeography* 4: 1. <https://doi.org/10.2307/3038124>
- Wheeler A, Jordan DR (1990) The status of the barbel, *Barbus barbus* (L.) (Teleostei, Cyprinidae), in the United Kingdom. *Journal of Fish Biology* 37: 393–399. <https://doi.org/10.1111/j.1095-8649.1990.tb05870.x>
- Wolter C, Röhr F (2010) Distribution history of non-native freshwater fish species in Germany: how invasive are they? *Journal of Applied Ichthyology* 26: 19–27.
- Vander Zanden MJ, Olden JD (2008) A management framework for preventing the secondary spread of aquatic invasive species. *Canadian Journal of Fisheries and Aquatic Sciences* 65: 1512–1522. <https://doi.org/10.1139/F08-099>